

Bisphenol TMC exhibits greater estrogenic activity than Bisphenol A and three other structural analogues exemplified by higher estrogen receptor α -mediated gene expression and breast cancer cell proliferation

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ARTICLE INFO

Keywords:

Bisphenol A
Bisphenol TMC
Endocrine disruptor
Estrogen receptor α
Regrettable substitution
Breast cancer

ABSTRACT

Bisphenol A (BPA) and its structural analogues are widely used in plastics production, raising concern due to endocrine-disrupting properties. While many analogues share structural similarities with BPA, their endocrine-disrupting effects remain insufficiently characterized. Cyclo-di-bisphenol A diglycidyl ether (cyclo-di-BADGE), tetrabromobisphenol S (TBBPS), bisphenol SIP (BPSIP), and bisphenol TMC (BPTMC) are particularly understudied. We assessed the estrogenic activity of these four BPA analogues compared to BPA. Transactivation assays in HEK-293 cells expressing estrogen receptor alpha (ER α) revealed that BPTMC was a more potent ER α agonist than BPA, with an EC₅₀ of 87 ± 20 nM versus 400 ± 100 nM for BPA, while the other tested analogues showed no significant agonistic activity. *In silico* analysis attributed this higher affinity to greater hydrophobicity and a bulkier bridging group between its phenolic rings. None of the compounds inhibited 17 β -hydroxysteroid dehydrogenase type 1 (17 β -HSD1) activity. However, BPTMC selectively inhibited 17 β -HSD2 (IC₅₀ = 4.8 ± 0.6 μ M) but not BPA. Importantly, 24 h exposure of ER α -positive MCF-7 breast cancer cells to 1 μ M BPTMC upregulated the expression of the ER α target genes *GREB1*, *TFF1*, and *PGR*, comparable to 10 nM E2, which was abolished by 100 nM of the ER α antagonist fulvestrant. Moreover, BPTMC stimulated MCF-7 cell proliferation at nanomolar concentrations over 72 h, and cell count analyses confirmed this effect. BPA also increased cell numbers, and both effects were reversed by fulvestrant. Collectively, we identified BPTMC as a potent ER α agonist capable of eliciting transcriptional and mitogenic responses at low concentrations, raising concerns about its endocrine-disrupting and breast cancer-promoting effects.

Abbreviations: ΔG_{bind} , Binding free energy; $\Delta G_{\text{desolvation}}$, Desolvation energy; 17 β -HSD1, 17 β -hydroxysteroid dehydrogenase type 1; 17 β -HSD2, 17 β -hydroxysteroid dehydrogenase type 2; ANOVA, Analysis of variance; BADGE, Bisphenol A diglycidyl ether; BPA, Bisphenol A; BPAF, Bisphenol AF; BPB, Bisphenol B; BPF, Bisphenol F; BPP, Bisphenol P; BPS, Bisphenol S; BPSIP, Bisphenol SIP; BPTMC, Bisphenol TMC; BPZ, Bisphenol Z; CSO, S-hydroxycysteine; Cyclo-di-BADGE, Cyclo-di-bisphenol A diglycidyl ether; DMEM, Dulbecco's Modified Eagle Medium; DPM, Diphenylmethane; DPP, Diphenylpropane; E1, Estrone; E2, 17 β -estradiol; EC₅₀, Half-maximal effective concentration; ECHA, European Chemicals Agency; EDCs, Endocrine-disrupting chemicals; ER α , Estrogen receptor alpha; ER β , Estrogen receptor beta; ERE, Estrogen response element; *GREB1*, Growth regulation by estrogen in breast cancer 1; HEK-293, Human embryonic kidney cells; IC₅₀, Half-maximal inhibitory concentration; LBP, Ligand-binding pocket; LogP, Lipophilicity; LogS, Solubility; MD, Molecular dynamics; MPP, Methylpiperidino pyrazole; NDGA, Nordihydroguaiaretic acid; PBTK, Physiologically-based toxicokinetic; PCOS, Polycystic ovary syndrome; PGR, Progesterone receptor; RMSF, Root mean square fluctuation; TBBPA, Tetrabromobisphenol A; TBBPS, Tetrabromobisphenol S; TCBPA, Tetrachlorobisphenol A; *TFF1*, Trefoil factor 1 also known as *pS2*; TMBPA, Tetramethylbisphenol A; TMBPF, Tetramethylbisphenol F.

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<https://doi.org/10.1016/j.tox.2025.154393>

Received 26 September 2025; Received in revised form 12 December 2025; Accepted 31 December 2025

Available online 31 December 2025

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1. Introduction

Endocrine-disrupting chemicals (EDCs) are a diverse group of exogenous compounds capable of interfering with hormonal systems, potentially leading to a range of adverse health effects in humans (Diamanti-Kandarakis et al. 2009). These chemicals can mimic or antagonize natural hormones, disrupt the synthesis, protein binding, transport, and metabolism of endogenous hormones, and alter the normal functioning of hormonal pathways (Hassan et al. 2024; Schug et al. 2011).

Among the most extensively studied EDCs is bisphenol A (BPA), a synthetic monomer widely used in the production of polycarbonate plastics and epoxy resins (Michalowicz, 2014). These materials are commonly found in consumer products such as food and beverage containers, dental sealants, medical devices, and thermal paper receipts (Hoekstra and Simoneau, 2013; Manzoor et al. 2022). The widespread use of BPA and its tendency to leach from products explain its presence in various human biological fluids, including blood, urine, and breast milk (Manzoor et al. 2022; Mercogliano and Santonicola, 2018). Notably, BPA has been detected in over 90 % of samples in the European and American population (Becker et al. 2009; Calafat et al. 2008; Mustieles et al. 2020). Increasing scientific and public concern has been directed toward the potential adverse health outcomes associated with chronic BPA exposure, including obesity, diabetes mellitus, developmental abnormalities, and hormone-sensitive cancers such as breast cancer (Cull and Winn, 2024; Duenas-Moreno et al. 2023; Rochester, 2013; Seachrist et al. 2016). In addition, rising infertility rates in developed countries prompt concerns regarding the potential impact of EDCs on female and male reproductive health (ovarian and Leydig cell steroidogenesis; follicle development, and sperm production) and fertility (Mlynarcikova et al. 2014; Odermatt et al. 2016). Hormone- and metabolism-related disorders such as polycystic ovary syndrome (PCOS) and endometriosis are increasingly recognized as susceptible to EDCs exposure (Bruni et al. 2022). Emerging evidence suggests a strong association between BPA exposure and the pathogenesis of PCOS, the most prevalent endocrine disorder among reproductive-aged women (March et al. 2010; Neuvonen et al. 2023; Srnovsni et al. 2023).

Several studies have demonstrated that many EDCs exert their effects by targeting estrogen receptor (ER) signaling. These chemicals can mimic or block the actions of endogenous estrogens, thereby disrupting ER-mediated transcriptional activity in both cellular and animal models (Diamanti-Kandarakis et al. 2009). Mechanistically, BPA can bind to and activate both ER α and ER β , triggering estrogenic signaling pathways and contributing to its endocrine-disrupting properties (Stanojevic and Sollner Dolenc, 2025; Welshons et al. 2006). ER α , in particular, belongs to the nuclear receptor superfamily and can be activated by ligands like 17 β -estradiol (E2), leading to its interaction with estrogen response elements (EREs) and subsequent transcription of target genes (Arnal et al. 2017). However, it can also be activated by other signaling pathways in a ligand-independent manner (Dero and Korach, 2006).

In addition to acting via ERs, BPA also affects estrogen homeostasis by altering steroid hormone biosynthesis. One key target group includes the hydroxysteroid dehydrogenases, particularly 17 β -hydroxysteroid dehydrogenase type 1 (17 β -HSD1) and type 2 (17 β -HSD2). 17 β -HSD1 catalyzes the conversion of estrone (E1), a weak estrogen, into E2, the most potent endogenous estrogen, thereby enhancing estrogenic activity (Hilborn et al. 2017). This enzyme is expressed in various estrogen-responsive tissues, including the breast, ovary, placenta, and endometrium (Miettinen et al. 1996; Poutanen et al. 1992a; Poutanen et al. 1992b; Soderqvist et al. 1998). BPA has been reported to inhibit 17 β -HSD1 with an IC₅₀ of approximately 113 μ M (Chen et al. 2023), a concentration unlikely to be reached in an exposure scenario *in vivo* and thus not considered to be toxicologically relevant. In contrast, 17 β -HSD2 catalyzes the oxidative conversion of E2 back to E1, thus reducing estrogenic potency, and is highly expressed in tissues such as the breast, uterus, prostate, liver, and intestine (Peltoketo et al. 1999). To date, no

data are available regarding the inhibition of 17 β -HSD2 by BPA or its analogues, underscoring the need for further investigation.

The growing evidence for adverse health effects of BPA has led to restrictions and bans limiting its use in consumer products, e.g., in Canada, Europe, and the USA beginning as early as 2008, which have been progressively extended from baby bottles to thermal paper receipts, and more recently, the complete ban in food contact materials (European Commission, 2024; Government of Canada, 2023; Kiss, 2013; Sieck et al. 2024). In response, the industry has introduced a variety of structural analogues such as bisphenol SIP (BPSIP, also referred to as D-8 or Wincon-8), bisphenol TMC (BPTMC), tetrabromobisphenol S (TBBPS), and bisphenol A diglycidyl ether (BADGE) as replacements for BPA. Cyclo-di-bisphenol A diglycidyl ether (Cyclo-di-BADGE), a cyclic product formed from BPA and BADGE during the production of epoxy resins, is a known contaminant in food contact materials. These derivatives share similar chemical structures and physicochemical properties with BPA, raising concerns that they may also exhibit endocrine-disrupting potential (See chemical structures in Fig. 1). Although manufacturers promote these chemicals as safer alternatives to BPA, scientists and regulators emphasize that further research is essential to assess their safety and understand their potential long-term effects on human health.

Given the urgent need for thorough toxicological and mechanistic evaluation of BPA analogues, we aimed to investigate the potential endocrine-disrupting activities of four BPA analogues, namely cyclo-di-BADGE, TBBPS, BPSIP, and BPTMC. Specifically, we assessed their ER α agonistic and antagonistic activities, along with their ability to inhibit 17 β -HSD1 and 17 β -HSD2. Additionally, *in silico* analysis was employed to investigate the underlying structural basis for estrogenic activity towards ER α . These endpoints were selected to provide a comparative understanding of the hormonal activity of BPA analogues relative to BPA itself and to inform risk assessment and regulatory decisions regarding their use.

2. Materials and methods

2.1. Chemicals and reagents

BPA (CAS# 80-05-7, Cat# 239658, purity \geq 99 %) was purchased from Merck (Darmstadt, Germany), cyclo-di-BADGE (CAS#

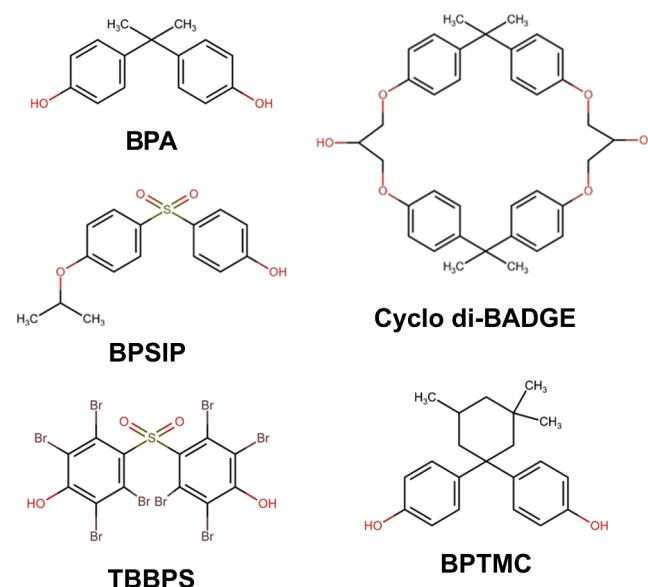


Fig. 1. Chemical structures of bisphenol A (BPA) and the four derivatives: cyclo-di-bisphenol A diglycidyl ether (cyclo-di-BADGE), tetrabromobisphenol S (TBBPS), bisphenol SIP (BPSIP), and bisphenol TMC (BPTMC).

20583-87-3, Cat# C10586.36, purity \geq 95 %) from Chiron AS (Trondheim, Norway), TBBPS (CAS# 39635-79-5, Cat# FT147403, purity \geq 95 %) from Biosynth AG (Staad, Switzerland), BPSIP (CAS# 95235-30-6, Cat# CBLQA-6028, purity \geq 98 %), and BPTMC (CAS# 129188-99-4, Cat# CBLQV-5232, purity \geq 98 %) from Chemie Brunschwig AG (Basel, Switzerland). Dimethyl sulfoxide (DMSO) was purchased from Huberlab (Aesch, Switzerland; CAS# 67-68-5, Cat# A3672). Fulvestrant (Merck; CAS# 129453-61-8, Cat# I4409) was a generous gift from Prof. Mohamed Bentires-Alj (Department of Biomedicine, University of Basel, Switzerland). Methylpiperidino pyrazole (MPP, CAS# 911295-24-4, Cat# M7068), apigenin (CAS# 520-36-5, Cat# A3145), nordihydroguaiaretic acid (NDGA, CAS# 500-38-9, Cat# 74540), E1 (CAS# 53-16-7, Cat# E9750), E2 (CAS# 50-28-2, Cat# E8875), and TritonTM X-100 (CAS# 9036-19-5, Cat# X100) were purchased from Merck. Stock solutions were prepared in DMSO and stored at -20°C .

Dulbecco's Modified Eagle Medium (DMEM, 4.5 g/L glucose, 2 mM L-glutamine), penicillin/streptomycin (10,000 U/mL and 10 mg/mL, respectively), 10 mM HEPES buffer, pH 7.4, and MEM non-essential amino acids were purchased from BioConcept (Allschwil, Switzerland). Fetal bovine serum (FBS, Cat# S1810; Batch S00CJ) was purchased from Biowest (Nuillé, France). The Tropix kit was obtained from Applied Biosystems (Foster City, CA, USA). [6,7-3H]-estrone was purchased from Anawa (Kloten, Switzerland) and [2,4,6,7-3H]-estradiol from PerkinElmer (Boston, MA, USA), and all other chemicals from Merck if not stated otherwise. The plasmids pRC/CMV-17 β HSD1 (originally from Dr. Reijo Vihko, University of Oulu, Finland) and pCMV6-17 β HSD2 were kindly provided by the late Dr. Stefan Andersson (University of Texas Southwestern Medical Center, Dallas, Texas, USA) (Khan et al. 2004; Wu et al. 1993).

2.2. Cell culture

Human Embryonic Kidney cells (HEK-293, ATCC, Manassas, VA, USA) were cultured in DMEM high glucose (4.5 g/L), 2 mM L-glutamine, supplemented with 10 % FBS, 10 mM HEPES, pH 7.4, 100 $\mu\text{g}/\text{mL}$ streptomycin and 100 U/mL penicillin, and 1 % (v/v) 100x MEM non-essential amino acids. MCF-7 cells (HTB-22, ATCC) were cultured in DMEM high glucose (4.5 g/L), 2 mM L-glutamine, supplemented with 10 % FBS and 10 mM HEPES, pH 7.4. Both cell lines were cultured under standard conditions (37°C , 5 % CO_2) in an incubator (Thermo Fisher Scientific, Waltham, USA). Passaging was carried out via trypsinization with 1x trypsin-EDTA (Merck; Cat# T4174; 1:10 diluted with PBS), and cell counts were determined using the Countess 3 FL automated cell counter (Thermo Fisher Scientific) with the trypan blue exclusion method.

2.3. ER α -dependent transactivation assay

HEK-293 cells (100,000 cells/well in 500 μL medium/well) were seeded in a 24-well plate (Sarstedt, Nümbrecht, Germany) coated with 0.005 % poly-L-lysine (0.01 % poly-L-lysine solution (Sigma-Aldrich, Steinheim, Germany) mixed 1:1 with PBS) and left to attach for 18 h. Plasmid DNA (390 ng/well for ER α ; 25 ng/well for CMV-LacZ; 390 ng/well for 4x ERE Luc) was diluted in 37°C water to a final volume of 30 μL /well. A volume of 10 μL /well of 37°C 2 M CaCl_2 was added slowly. The mixture was pipetted dropwise into an equal volume of 37°C 2x BEST solution (8.0 g NaCl; 0.198 g $\text{Na}_2\text{HPO}_4 \cdot 7 \text{H}_2\text{O}$; 5.3 g N,N-bis[2-hydroxyethyl]-2-aminoethane sulfonic acid (Sigma-Aldrich), add water to 500 mL, pH 7.0) and incubated at room temperature (RT) for 15 min. The cells were transfected by adding 80 μL of this transfection mixture to the medium of each well. After a period of 4 h, the supernatant in each well was removed and replaced with fresh medium (500 μL /well). The treatment medium was phenol-red free DMEM high glucose (4.5 g/L) (Sigma-Aldrich) supplemented with L-glutamine, 10 mM HEPES, pH 7.4 (BioConcept), 100 $\mu\text{g}/\text{mL}$ streptomycin and 100 U/mL penicillin, and

1 % (v/v) 100x MEM non-essential amino acids (BioConcept) as well as 10 % charcoal-treated FBS. The charcoal treatment was performed to deprive the FBS of E2. For charcoal treatment, 50 mL FBS were mixed with 2 g of activated charcoal (Sigma-Aldrich) and shaken at 4°C overnight. The suspension was subjected to centrifugation at 4°C and 15,000 g for 10 min, after which the supernatant was decanted into a new tube. The aforementioned procedure was repeated for a total of two overnight charcoal treatments. Following the final charcoal treatment, the FBS was sterile filtered using a 0.2 μm Filtrpur S 0.2 filter (Sarstedt). One h prior to the treatment, the medium was replaced with the E2-deprived medium. Subsequently, the cell supernatant was replaced with 500 μL of the same medium containing the test compounds at their respective concentrations and a fixed amount of 0.1 % DMSO. To investigate possible antagonistic effects of test compounds on ER α , 20 $\mu\text{L}/\text{well}$ of 52 nM E2 was added after another hour, resulting in a total concentration of 2 nM E2 in the well. For the assessment of agonistic and antagonistic effects of BPA and the four analogues, cells were examined morphologically under a light microscope throughout the entire treatment period. No signs of cytotoxicity or stress-related changes (e.g., cell rounding, detachment, or granulation) were observed at any of the tested concentrations. After 24 h of treatment, the cell supernatant was removed, and the cells were lysed for a period of 5 min at RT with 60 μL of Tropix lysis solution (Applied Biosystems). The cell lysate was stored at -80°C until further processing. Luciferase activity was determined by adding 100 μL of a D-luciferin substrate solution into 20 μL of the lysate in pure grade white 96-well microtiter plates (BRAND, Wertheim, Germany). The luciferin solution consisted of 0.56 mM D-luciferin, 63 mM ATP, 0.27 mM CoA, 0.13 mM EDTA, 33.3 mM dithiothreitol, and 8 mM MgSO₄. The luminescence was then measured with the SpectraMax L Microplate Reader (Molecular Devices, San Jose, CA, USA). The respective β -galactosidase activity was quantified with the same plates and device using 20 μL of the lysate and the Galacto-Light PlusTM kit (Applied Biosystems). The luciferase activity of each well was normalized to the respective β -galactosidase activity.

2.4. Computational modeling

The crystal structure of the ER α (PDB ID 4MGD) (Delfosse et al. 2014) was downloaded and pre-processed for further analyses assuming pH 7.4 ± 0.5 using the protein preparation wizard in Schrödinger (version 2024-3; Protein Preparation Workflow; Epik; Impact; Prime; Schrödinger, LLC, New York, NY, 2024). This crystal structure was chosen based on its high resolution (1.9 \AA), specifically around the ligand HPTE (bisphenol-class compound) featuring a bulky trifluoromethyl group. This structure presumably provides a more suitable receptor conformation for accommodating the voluminous linker of BPTMC than the crystal structure with bound BPA (PDB ID 3UU7). Missing loops were filled with Prime, and protonation states were calculated with Epik. S-hydroxycysteine (CSO) residues were mutated back to cysteines, and Ser537, a mutation introduced for crystallization, was mutated back to the original residue Tyr537. To obtain a complex with BPA, Glide docking (standard precision) was used (Friesner et al. 2004). To obtain a complex with BPTMC, Dolina flexible docking was employed, due to the larger size of BPTMC compared to HPTE (Smiesko, 2013). The best ranked poses were selected for further processing.

The protein-ligand complexes were neutralized by adding counterions and solvated in TIP3P water (orthorhombic box, 10 \AA). A salt concentration of 0.15 M was added and minimization was carried out using the OPLS 2005 force field. The system was equilibrated using the standard protocol of Desmond (version 2023-4; Desmond Molecular Dynamics System, D. E. Shaw Research, New York, NY, 2023. Maestro-Desmond Interoperability Tools, Schrödinger, New York, NY, 2023). Then, three replicate molecular dynamics (MD) simulations of 50 ns were performed for both complexes, using the default settings of Desmond (version 2023-4). In total, 1000 frames were written out for each simulation. Ligand root mean square fluctuation (RMSF) was calculated

for the atoms of the bisphenol core structure (excluding substituents on the bridge). The RMSF reflects the standard deviation in specified atom position during the MD trajectory and is calculated following Eq. 1 (Tsukamoto et al. 2024):

$$RMSF = \sqrt{\langle (r_i - \langle r_i \rangle)^2 \rangle} \quad (1)$$

where r_i is the position of atom i and $\langle r_i \rangle$ is the ensemble average position of atom i . In this case, a high RMSF value indicates large fluctuations in the position of the bisphenol core atoms during the simulation, which means more movement of the bisphenol core in the binding pocket of ER α .

For every MD simulation, the binding free energy was calculated with thermal MM/GBSA script (Schrödinger, version 2024–3). For the calculation, frames 100–500 were selected to allow for a short equilibration period, and a step size of 10 was used. From the MM/GBSA data, the desolvation energy was extracted by taking the negative of the solvation energy. Values were reported as the mean \pm SD. Lastly, Qikprop (Version 2024–3; QikProp, Schrödinger, LLC, New York, NY, 2024) was used to calculate basic physico-chemical descriptors of BPA and BPTMC.

2.5. 17β -HSD1 activity assay

The 17β -HSD1 activity was determined as described previously, with some modifications (Engeli et al. 2017). Briefly, 2×10^6 HEK-293 cells were seeded per 10 cm dish in 10 mL medium, left to attach for 18 h and transiently transfected with 8 μ g of plasmid for human 17β -HSD1 using a modified version of the calcium phosphate precipitation method (Graham and van der Eb, 1973). The plasmid DNA was diluted in 37°C water to a final volume of 437.5 μ L/dish. A volume of 62.5 μ L/dish of 37°C 2 M CaCl_2 was added slowly. The mixture was pipetted dropwise into an equal volume of 37°C 2x BEST solution (described above in Section 2.3) and incubated at RT for 15 min. The cells were transfected by adding 1 mL/dish of the prepared DNA- CaCl_2 -BEST transfection mixture to the culture medium. After a period of 4 h the medium containing the transfection mix was removed and replaced with fresh medium (10 mL/dish). After a 48 h period following transfection, cells were collected in 2 mL ice cold PBS, then portioned in 200 μ L aliquots and pelleted for 4 min at 13,000 g at 4°C prior to snap freezing in liquid nitrogen and storage at -80°C . For each experiment, cell pellets were resuspended in 1 mL TS2 buffer (100 mM NaCl, 1 mM EGTA, 1 mM EDTA, 1 mM MgCl_2 , 250 mM sucrose, 20 mM Tris HCl, pH 7.4) and lysed by sonication (UP50H sonicator, Hieltscher Ultrasonics). Cell lysates were diluted to obtain 30 % substrate to product conversion after a 10 min incubation. The diluted HEK-293 cell lysate (10 μ L) containing overexpressed 17β -HSD1 was pre-incubated with either DMSO as solvent control or DMSO-dissolved test compound diluted in TS2 buffer in a total volume of 12.2 μ L for 8 min on ice. The total solvent concentration in the reaction was 0.1 %. The reaction was started by adding 10 μ L of a reaction mixture containing substrate and co-substrate. 17β -HSD1 activity was determined in the presence of 200 nM E1 in TS2 buffer, including 81 pCi/ μ L of [6,7-3H]-estrone (Anawa) and 454 μ M NADPH. The reaction mixture was incubated for 10 min at 37°C, after which the reaction was terminated with 10 μ L pure methanol containing 5 mM E1 and 5 mM E2. 10 μ L of each sample was applied to TLC plates (Macherey Nagel, Oensingen, Switzerland), and steroids were separated using a mobile phase consisting of chloroform and ethyl acetate in a 4:1 ratio. The plates were dried overnight, bands of E1 and E2 were visualized under UV light, marked, excised, and analyzed via scintillation counting using a Tri-Carb 2900TR liquid scintillation analyzer (Packard, Meriden, Connecticut, USA) with the IRGASAFE Plus scintillation cocktail (Zinsser Analytic, Eschborn, Germany). To determine the enzymatic activity of 17β -HSD1, the conversion rate of E1 to E2 was calculated and normalized to the DMSO control.

2.6. 17β -HSD2 activity assay

17β -HSD2 activity was determined as described previously, with some modifications (Engeli et al. 2017). Briefly, 2×10^6 HEK-293 cells were seeded per 10 cm dish in 10 mL medium, left to attach for 18 h and transiently transfected with 8 μ g of plasmid for human 17β -HSD2 as described above for 17β -HSD1, followed by preparation of cell pellets that were stored at -80°C . For activity assays, cell pellets were resuspended in 1 mL TS2 buffer and diluted to obtain 30 % conversion. The diluted HEK-293 cell lysate (10 μ L) containing overexpressed 17β -HSD2 was pre-incubated with either DMSO as solvent control or DMSO-dissolved test compound diluted in TS2 buffer in a total volume of 12.2 μ L for 8 min on ice. The total solvent concentration in the reaction was 0.1 %. The reaction was started by adding 10 μ L of the substrate and co-substrate. 17β -HSD2 activity was determined in the presence of 200 nM E2 in TS2 buffer, including 3 Bq/ μ L (81 pCi/ μ L) of [2,4,6,7-3H]-estradiol (PerkinElmer) and 454 μ M NAD $^+$. The reaction mixture was incubated for 10 min at 37°C, followed by termination of the reaction by adding 10 μ L methanol containing 5 mM E1 and 5 mM E2. 10 μ L of each sample was applied to TLC plates (Macherey Nagel), and steroids were separated using a mobile phase consisting of chloroform and ethyl acetate in a 4:1 ratio. The plates were dried overnight, bands of E1 and E2 were visualized under UV light, marked, excised, and analyzed via scintillation counting using a Tri-Carb 2900TR liquid scintillation analyzer (Packard) with the IRGASAFE Plus scintillation cocktail (Zinsser Analytic). To determine the enzymatic activity of 17β -HSD2, the conversion rate of E2 to E1 was calculated and normalized to the DMSO control.

2.7. Quantification of mRNA by RT-qPCR

MCF-7 cells were seeded in a 6-well plate in normal culture medium at a density of 450,000 cells/well and allowed to attach overnight. To deprive the cells of steroids, the medium was replaced 4 h prior to treatment with phenol-red-free DMEM with high glucose (4.5 g/L) (Sigma-Aldrich), 10 mM HEPES, pH 7.4 (BioConcept), supplemented with 10 % FBS (Biowest) that was charcoal-treated as described above (Section 2.3). Treatment was initiated by replacing the steroid-free medium with the treatment medium which consisted of the same steroid-deprived medium but contained the compounds of interest at the desired concentration as well as 0.1 % DMSO. Following a 24 h treatment period, the cells were washed with ice-cold PBS and then frozen at -80°C . MCF-7 cells were lysed with 350 μ L of RLT buffer (Qiagen, Hombrechtikon, Switzerland) and the cell lysate was transferred to RNeasy® Plus Mini gDNA eliminator columns and then centrifuged for 1 min at 13,000 g. From the eluate, total RNA was extracted according to the manufacturer (Qiagen RNeasy® Plus Mini Extraction kit). We assessed the RNA concentration and quality using a NanoDrop 2000 (Thermo Scientific). cDNA was synthesized from 1 μ g RNA using PrimeScript RT Reagent Kit (Takara Bio Europe SAS; Saint-Germain-en-Laye, France) according to the manufacturer. The real-time PCR measurement of individual cDNAs was performed in triplicate using SYBR green dye (Roche Diagnostics, Rotkreuz, Switzerland) containing 10 μ M of each primer (sense and antisense) (see Suppl. Table 1). Relative mRNA levels were calculated using the $\Delta\Delta\text{C}_t$ method with *Peptidylprolyl isomerase A (PPIA)* as housekeeping gene (Tsachaki et al. 2020).

2.8. Real-time cell proliferation assay

MCF-7 cells were seeded in a 96-well plate in normal culture medium at a density of 5,000 cells per well and allowed to attach overnight. Cells were treated in phenol-red-free DMEM with high glucose (4.5 g/L; Sigma-Aldrich) supplemented with 10 % steroid-depleted FBS (Biowest). For the steroid depletion, the FBS was charcoal-treated as described in Section 2.3. To deprive the cells of steroids before the start of the treatment, medium was replaced with the steroid-deprived

medium 3 h prior to treatment. The 3 h steroid deprivation was optimized to reduce background estrogenic activity and sensitize cells to E2, while maintaining cell viability and avoiding stress from extended culture in charcoal-stripped, phenol red-free medium. The concentration of DMSO did not exceed 0.1 % in the treatment medium and was maintained at a constant level across all treatments. After starting the treatment, the cells were monitored for 72 h with the Celloger Mini Plus™ (Curiosis, Seoul, South Korea). The confluence of the cells was determined with the CellogerAnalysis App Version 2.1.2023.0512 (Curiosis). The doubling time was calculated with a non-linear fit (exponential growth equation) in GraphPad Prism 10.2.0. (GraphPad Software, La Jolla, CA, USA). Each treatment was measured in three biological replicates with six technical replicates each.

2.9. Statistical analysis

Data are given as mean \pm SD of at least three independent experiments. Statistical analyses were performed using the GraphPad Prism 10.2.0. One-way analysis of variance (ANOVA) was then used for comparisons between groups, followed by Tukey or Dunnett's post-test procedure to localize differences between treatments containing chemicals and the vehicle group. Concentration response curves for estimation of half-maximal inhibitory concentration (IC_{50}) values were fitted and analyzed by non-linear regression. Statistical significance was set at $* p < 0.05$.

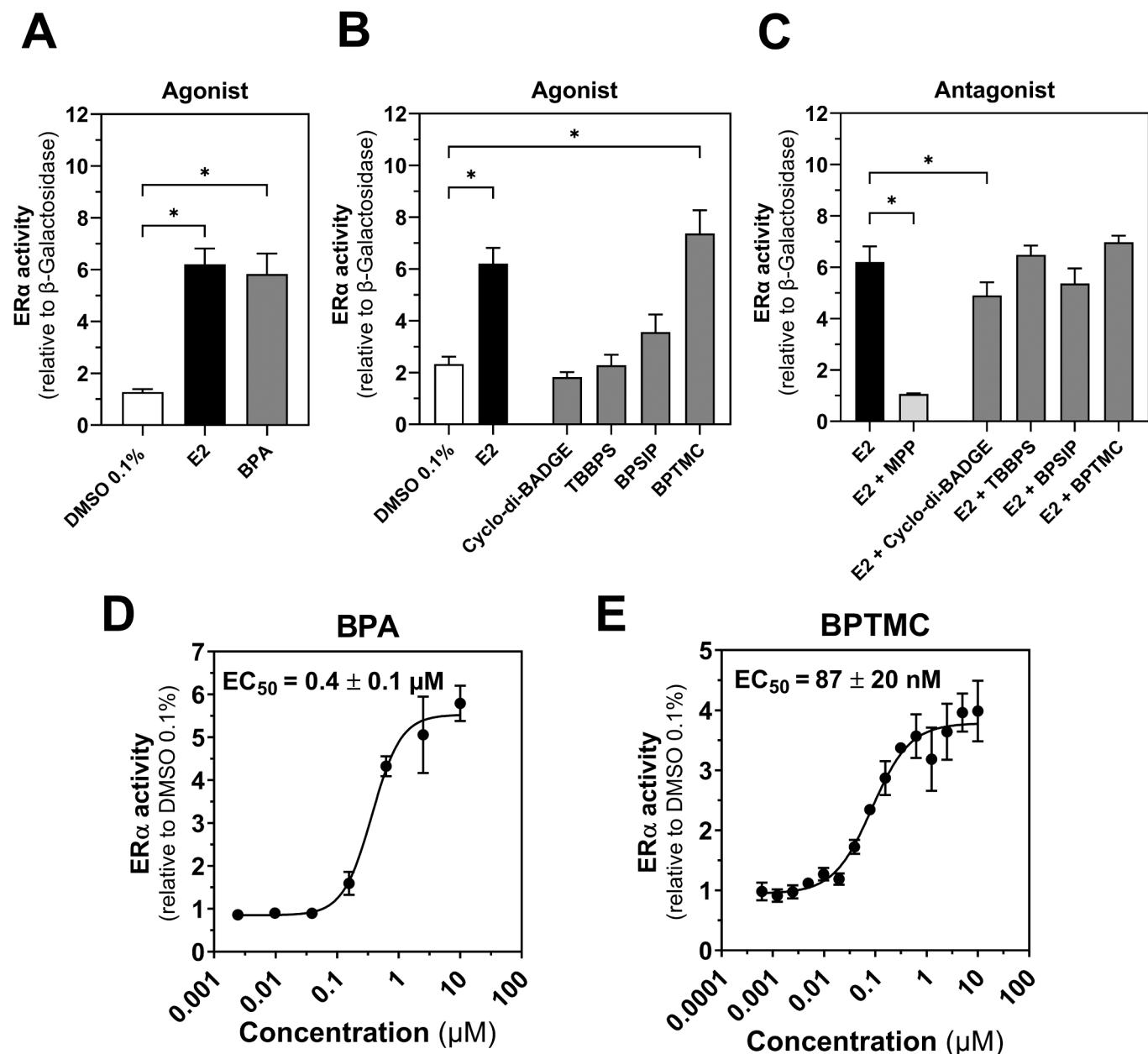


Fig. 2. Screening of BPA and its analogues for ER α agonistic and antagonistic activity. (A) ER α agonistic activity of 10 μ M BPA. E2 was used at 2 nM as positive control. (B) Screening of the BPA analogues cyclo-di-BADGE, TBBPS, BPSIP, and BPTMC at 10 μ M for ER α agonistic effects. (C) Screening of the BPA analogues cyclo-di-BADGE, TBBPS, BPSIP, and BPTMC at 10 μ M for ER α antagonistic effects in the presence of 2 nM E2. The selective ER α antagonist MPP was used at 2.5 μ M as a positive control. (D) EC_{50} curve of BPA on ER α agonistic activity. (E) EC_{50} curve of BPTMC on ER α agonistic activity. Values represent mean \pm SD of three independent experiments, each performed in technical duplicates. Statistical significance was determined using one-way ANOVA, followed by Dunnett's post-test with $*p < 0.05$.

3. Results

3.1. BPTMC is a significantly more potent ER α activator than BPA

To assess the estrogenic activity of BPA and its four analogues, we conducted luciferase reporter gene measurements using HEK-293 cells transiently expressing human ER α , a responsive luciferase construct and β -galactosidase as transfection control for normalization. Each compound was first tested at a concentration of 10 μ M to evaluate both agonistic and antagonistic effects on ER α , using a cell-based transactivation assay. As shown in Fig. 2A, BPA activated ER α in a clear agonistic response, confirming the findings of earlier studies (Gould et al. 1998). Among the analogues, BPTMC also induced ER α -dependent transactivation to a level comparable to the 2 nM E2 positive control, whereas cyclo-di-BADGE, TBBPS, and BPSIP showed no significant agonistic activity (Fig. 2B). In the presence of 2 nM E2, MPP at 2.5 μ M was used as positive control for a selective ER α antagonist. Our screening revealed that cyclo-di-BADGE at 10 μ M exhibited weak antagonistic activity (Fig. 2C). However, no concentration-dependent effect was observed in a follow-up quantification (data not shown). Notably, while BPTMC functioned as an ER α agonist, it did not further enhance the reporter gene activity in the presence of 2 nM E2, suggesting maximal activation of the reporter gene (Fig. 2C). For BPA, no relevant antagonistic effect on ER α has been observed at concentrations below 100 μ M (Delfosse et al. 2012; Durcik et al. 2022; Lee et al. 2023); therefore, it was not further investigated in our study. To quantify the relative potencies of BPA and BPTMC as ER α agonists, we conducted concentration-response experiments and determined their half-maximal effective concentrations (EC₅₀). BPA showed an EC₅₀ of $0.4 \pm 0.1 \mu$ M (Fig. 2D), while BPTMC demonstrated a significantly lower EC₅₀ of 87 ± 20 nM (Fig. 2E). Our findings indicate that BPTMC is significantly more potent than BPA in activating ER α .

3.2. Increased binding stability and lipophilicity enhance BPTMC's ER α interaction

To investigate how subtle structural differences between BPA and BPTMC affect their binding affinity to ER α , their physicochemical properties and binding poses were compared. As shown in Table 1, BPTMC showed a higher lipophilicity (logP), lower solubility (logS), and a smaller dipole moment than BPA, indicating greater overall hydrophobicity. This was further supported by the calculated desolvation energies ($\Delta G_{\text{desolvation}}$): BPTMC presented a 3.8 kcal/mol lower desolvation cost than BPA, suggesting a more favorable transfer from the solvent to the lipophilic ER α binding pocket. These results indicate that increased hydrophobicity of BPTMC contributes to its enhanced binding affinity for ER α .

Next, the stability of BPA and BPTMC during MD simulations was compared. Fig. 3A shows the initial binding pose of BPA at the start of the simulation. Notably, all crystal docked BPA poses formed interactions with Glu353 and Thr347. However, in the published crystal structure of BPA to ER α (PDB ID 3UU7; (Delfosse et al. 2012)), the compound is oriented with its bridge directed away from Phe404 and interacts instead with the thioether acceptor of Met421 rather than

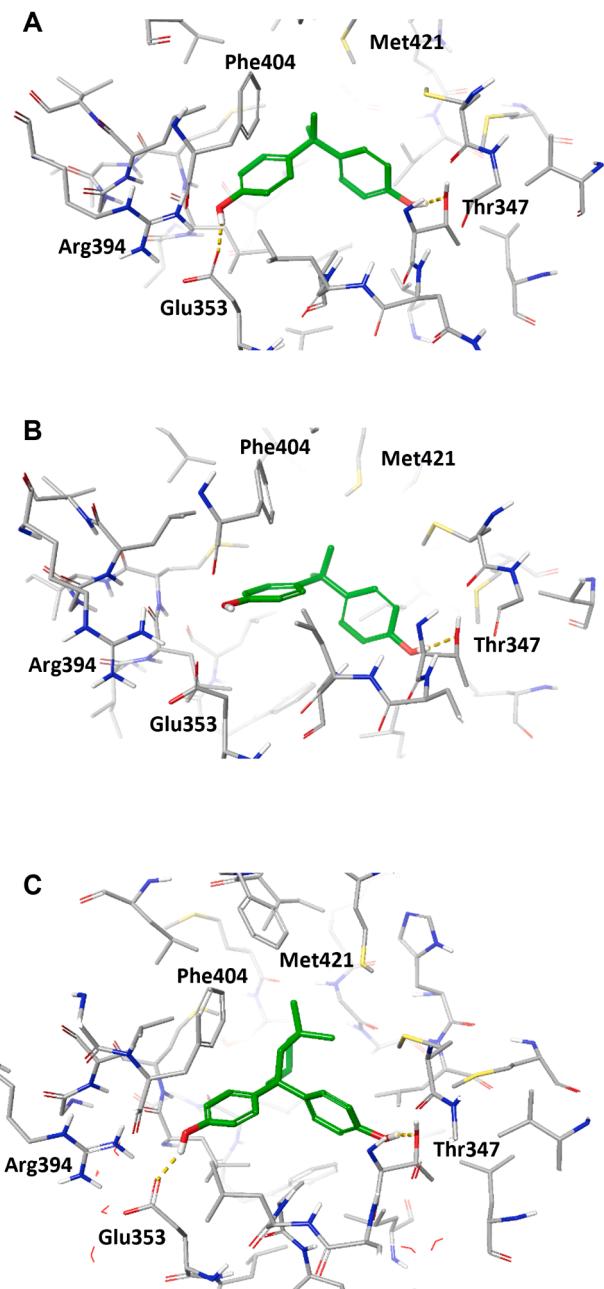


Fig. 3. Molecular dynamics of BPA and BPTMC binding to ER α . (A) BPA (green) bound to ER α at the start of MD simulations, forming hydrogen bonds (yellow dashed lines) with Glu353 and Thr347. (B) BPA (green) during MD simulations, losing interaction with Glu353. (C) BPTMC (green) bound to ER α , forming hydrogen bonds with Glu353 and Thr347. Images were generated in Schrödinger Maestro (2024-3).

hydroxyl oxygen of Thr347 (Suppl. Figure 1). The electron density around His524 in 3UU7 is weak, while some density is observed near Thr347, suggesting that BPA may adopt two alternative binding poses. In all three MD simulations, the interaction with Thr347 was consistently maintained. However, the interaction with Glu353 was repeatedly lost (Fig. 3B).

BPTMC adopted a similar binding pose to BPA (Fig. 3C). Unlike BPA, BPTMC maintained stable interactions with both Thr347 and Glu353 throughout the simulation. This increased stability is likely due to the bulkier substituent on its bridge, which limits ligand flexibility and mobility, thus anchoring the molecule firmly within the ER α binding pocket. The calculated average RMSF confirmed this higher stability of

Table 1

Property overview (calculated with QikProp) and desolvation energy (average calculated from three replicate MD simulations with MM/GBSA) for BPA and BPTMC.

Compound	BPA	BPTMC
LogP	3.3	4.5
LogS	-3.4	-5.6
Dipole moment [Debye]	2.1	1.2
ΔG_{bind} (kcal/mol)	-50.2 ± 3.5	-65.2 ± 5.5
$\Delta G_{\text{desolvation}}$ (kcal/mol)	-18.2 ± 1.3	-22.0 ± 2.0

the bisphenol core, with an average of $0.94 \pm 0.06 \text{ \AA}$ for BPTMC compared to $1.19 \pm 0.07 \text{ \AA}$ for BPA. Additionally, the larger ring structure of BPTMC enabled more extensive van der Waals interactions with surrounding lipophilic residues (Met343, Met421, Leu525, Leu540), whereas BPA did not fully exploit the lipophilic potential of the available cavity. The lower stability of BPA in the ER α binding pocket was further reflected by its higher binding free energy (ΔG_{bind}) compared to BPTMC, as shown in Table 1.

3.3. BPTMC, but not BPA, shows moderate inhibition of 17 β -HSD2

To investigate whether the BPA derivatives might interfere with the enzymes controlling the intracellular concentrations of active estrogens, we assessed their ability to inhibit the key steroid-converting enzymes 17 β -HSD1 and 17 β -HSD2. Initial screening assays were conducted at a compound concentration of 10 μM to evaluate potential enzyme inhibition (Figs. 4A and 4B). Compounds leading to less than 30 % residual enzymatic activity were selected for determining the IC₅₀. As expected, the positive controls apigenin and NDGA (Engeli et al. 2017) used at a concentration of 20 μM resulted in a significant inhibition of 17 β -HSD1 and 17 β -HSD2 activities, respectively (Figs. 4A and 4B). Interestingly, BPA at 10 μM did not alter the activity of either 17 β -HSD1 or 17 β -HSD2. In addition, none of the tested BPA analogues significantly inhibited 17 β -HSD1 activity, although BPTMC showed a tendency towards lowering its activity (Fig. 4A). Cyclo-di-BADGE, TBBPS, and BPSIP did also not elicit substantial inhibition of 17 β -HSD2 activity (Fig. 4B). Although TBBPS and BPSIP caused a statistically significant decrease in 17 β -HSD2 activity, the residual enzymatic activity remained above 30 % (Fig. 4B). In contrast, BPTMC markedly inhibited 17 β -HSD2, reducing enzyme activity to below 20 % at a concentration of 10 μM . The concentration-response analysis of BPTMC revealed a moderate inhibitory potency with an IC₅₀ value of $4.8 \pm 0.6 \mu\text{M}$ (Fig. 4C).

3.4. BPTMC upregulates the expression of ER α downstream target genes in MCF-7 breast cancer cells

To assess the transcriptional activity of ER α in a pathophysiologically relevant cellular context, we employed MCF-7 cells, a well-established, ER α -positive human breast cancer cell line. MCF-7 cells are widely used as a model system for studying estrogenic signaling due to their endogenous expression of functional ER α and their sensitivity to E2 (Hamelers et al. 2003). In particular, ER α activation in MCF-7 cells has been shown to induce the expression of canonically estrogen-responsive genes such as *Growth regulation by estrogen in breast cancer 1* (*GREB1*), *Trefoil factor 1*, also known as *pS2* (*TFF1*), and *Progesterone receptor* (*PGR*), all of which are established transcriptional targets of ER α (Duijndam et al. 2021).

Cell viability was first assessed using the colorimetric XTT assay. As shown in Suppl. Figure 2, the positive control (0.1 % Triton-X) decreased viability of MCF-7 cells. Notably, we found that viability was not significantly affected by exposure to BPTMC for 24 h up to 3 μM (Suppl. Figure 2).

To evaluate whether BPTMC can upregulate the expression of ER α target genes, MCF-7 cells were exposed to increasing concentrations (8 nM to 1 μM) of BPTMC, and mRNA levels of *GREB1* (Fig. 5A), *TFF1* (Fig. 5B), and *PGR* (Fig. 5C) were quantified by RT-qPCR. Treatment with 1 μM BPTMC for 24 h resulted in a marked upregulation of all three ER α -regulated genes, with expression levels comparable to those observed following exposure to the positive control 10 nM E2. Notably, an increase in *TFF1* mRNA levels was detectable at concentrations as low as 200 nM BPTMC, indicating a concentration-dependent estrogenic response. To confirm that the transcriptional activation was specifically mediated via ER α , co-treatment experiments were conducted using the selective ER α antagonist fulvestrant at 100 nM. Co-administration of fulvestrant with either 10 nM E2 or 1 μM BPTMC substantially attenuated the induction of the expression of *GREB1*, *TFF1*, and *PGR*,

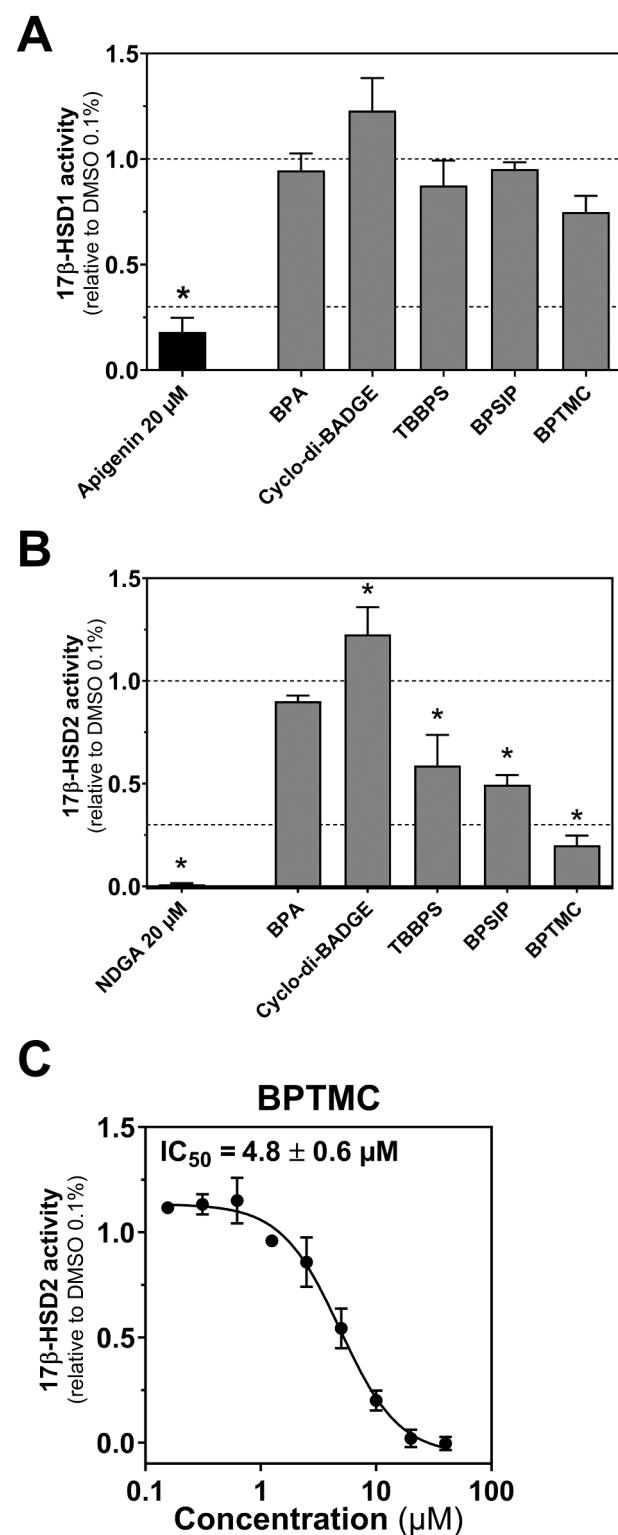


Fig. 4. Screening for 17 β -HSD1 and 17 β -HSD2 inhibitory activity by BPA and its analogues. (A) 17 β -HSD1 inhibitory activity of 10 μM BPA, cyclo-di-BADGE, TBBPS, BPSIP, and BPTMC was measured in lysates of 17 β -HSD1-transfected HEK-293 cells. 20 μM apigenin served as positive control for inhibition. (B) 17 β -HSD2 inhibitory activity of 10 μM BPA, cyclo-di-BADGE, TBBPS, BPSIP, and BPTMC was measured in lysates of 17 β -HSD2 transfected HEK-293 cells. 20 μM NDGA served as positive control for inhibition. (C) IC₅₀ curve for 17 β -HSD2 inhibition by BPTMC. An IC₅₀ value of $4.8 \pm 0.6 \mu\text{M}$ was obtained. Values represent mean \pm SD of three independent experiments, each performed in technical duplicates. Statistical significance was determined using one-way ANOVA, followed by Dunnett's post-test, with * $p < 0.05$.

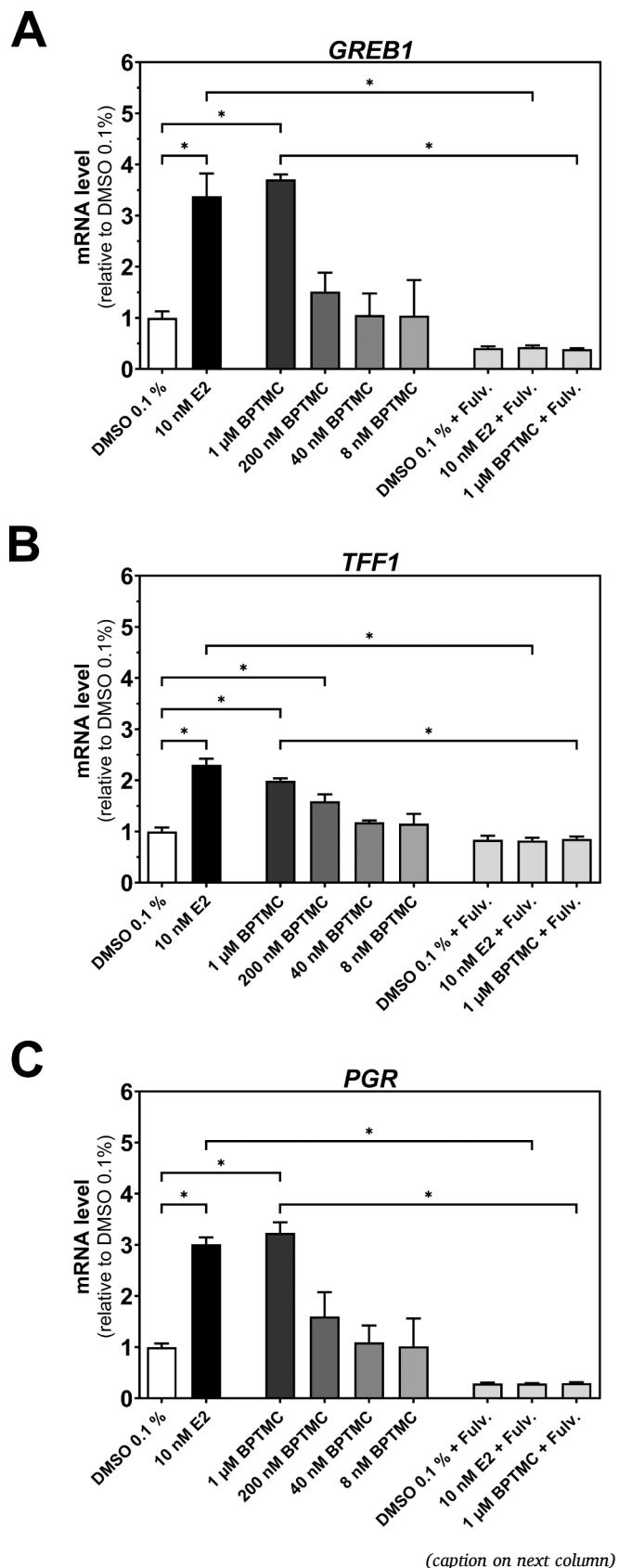


Fig. 5. mRNA levels of three downstream targets of ER α in MCF-7 breast cancer cells after exposure to various concentrations of BPTMC for 24 h. (A) mRNA levels of *GREB1* after 24 h exposure to the indicated concentrations of BPTMC. 10 nM E2 was used as a positive control for ER α activation. The increase in *GREB1* mRNA levels induced by E2 or BPTMC exposure was blocked by 100 nM of the selective ER α antagonist fulvestrant. (B) mRNA levels of *TFF1* after 24 h exposure to various concentrations of BPTMC. 10 nM E2 was used as a positive control for ER α activation. The increase in *TFF1* mRNA levels induced by E2 or BPTMC exposure was blocked by 100 nM fulvestrant. (C) mRNA levels of *PGR* after 24 h exposure to various concentrations of BPTMC. 10 nM E2 was used as a positive control for ER α activation. The increase in *PGR* mRNA levels induced by E2 or BPTMC exposure was blocked by 100 nM fulvestrant. Values represent mean \pm SD of three independent experiments, each performed in technical triplicates. Statistical significance was determined using one-way ANOVA, followed by Tukey's post-test, with * p < 0.05.

confirming that the effects of BPTMC were ER α -dependent (Figs. 5A, 5B, and 5C). These results demonstrate that BPTMC can upregulate estrogenic gene expression in ER α -positive breast cancer cells, supporting its classification as an ER α agonist.

3.5. Low nanomolar concentrations of BPTMC promote proliferation of MCF-7 breast cancer cells

To evaluate the proliferative effects of BPTMC on ER α -positive cells, we monitored MCF-7 breast cancer cell proliferation, following treatment with various concentrations of BPTMC, 10 nM E2 as positive control, or vehicle negative control. Cell proliferation was assessed over a 72 h period by measuring confluence (Fig. 6) and impedance (Suppl. Figure 3) in the steroid-deprived treatment medium, both of which are established indicators of cell growth and density. The endpoint of the cell proliferation was further validated by performing a cell count after the 72-h treatment period (Suppl. Figure 4). MCF-7 cells treated with vehicle control showed minimal proliferation, with confluence plateauing between 20 and 30 h post-treatment (Fig. 6A). In contrast, treatment with 10 nM E2 (Figs. 6B), 1 μ M BPTMC (Fig. 6C), or 100 nM BPTMC (Fig. 6D) resulted in sustained cell proliferation over the full 72 h period, culminating in significantly increased confluence and cell count compared to vehicle-treated cells (Fig. 6I and Suppl. Figure 4). These findings indicate that BPTMC stimulates MCF-7 cell proliferation in a concentration-dependent manner, with 100 nM being sufficient to elicit a significant mitogenic effect. These results were confirmed in impedance-based monitoring (Suppl. Figure 3). Notably, 72 h exposure at 100 nM resulted in a greater increase in cell number with BPTMC than with BPA (Suppl. Figure 4). At 10 nM BPTMC, no clear increase in confluence and cell number was observed (Fig. 6E and Suppl. Figure 4), suggesting that this concentration may be below the threshold required to drive robust cell proliferation. However, impedance-based monitoring revealed more ambiguous results at 10 nM (Suppl. Figures 3E and 3I), potentially indicating low-level or context-dependent proliferative activity not captured by assessing confluence and cell count. To confirm that the observed proliferative effects were indeed mediated by ER α activation, co-treatment experiments were performed using 100 nM of the selective ER α antagonist fulvestrant. Co-treatment with fulvestrant abolished the proliferative response to 10 nM E2 (Fig. 6G and Suppl. Figure 4), 1 μ M BPTMC (Fig. 6H and Suppl. Figure 4), and 1 μ M BPA (Suppl. Figure 4). Confluence and cell number in these groups remained comparable to those of the vehicle and fulvestrant control group (Fig. 6F and Suppl. Figure 4). The impedance-based monitoring confirmed that BPTMC-induced MCF-7 cell proliferation is ER α -dependent (Suppl. Figure 3).

As a next step, the corresponding doubling time for each condition was calculated (Suppl. Figure 5). The doubling time of control MCF-7 cells was 50.9 ± 8.1 h, which aligns well with the literature (Ruohola et al. 2001). Exposure to 10 nM E2 reduced the doubling time by approximately two-fold, consistent with the well-known

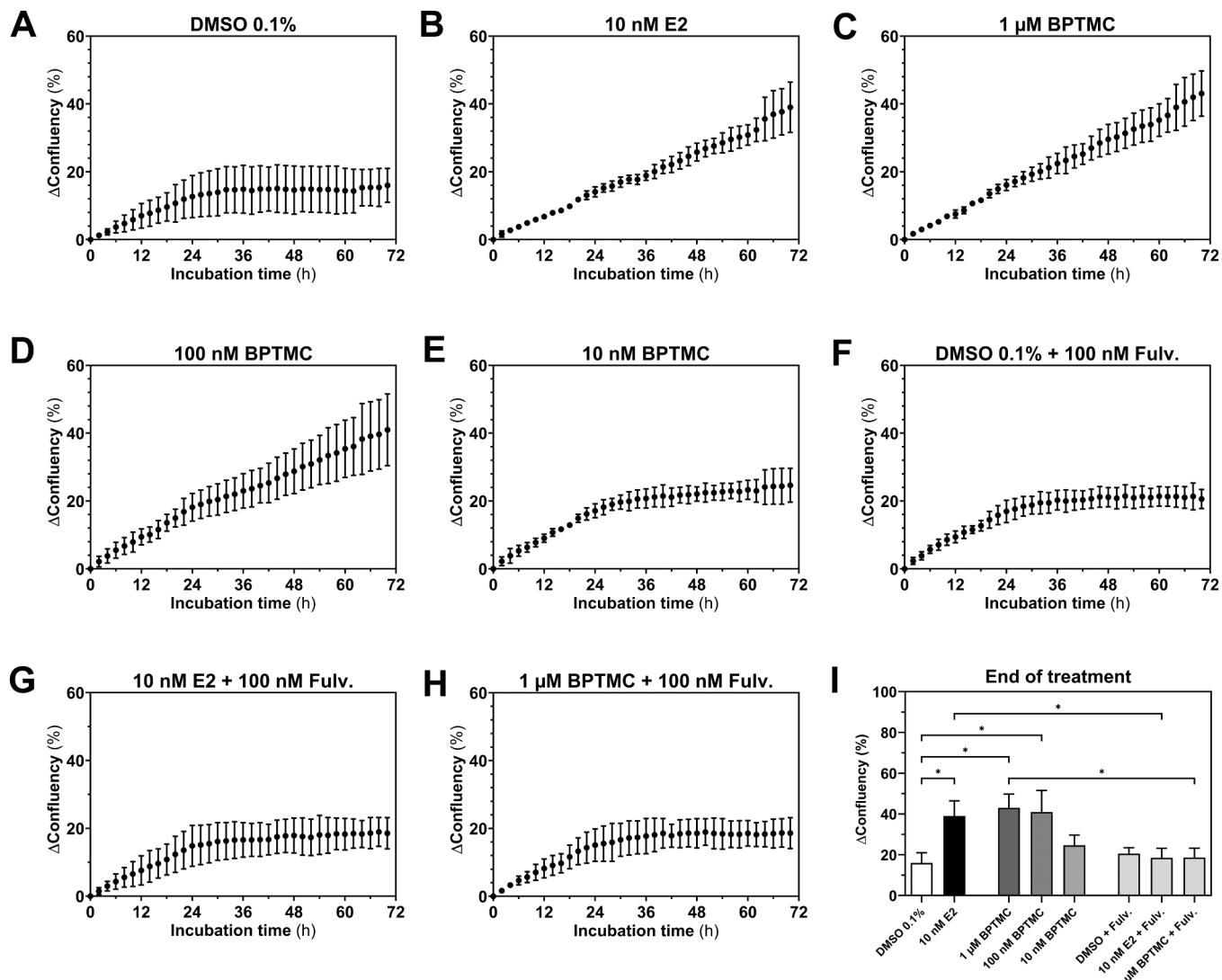


Fig. 6. MCF-7 cell proliferation in response to BPTMC. The difference in confluence (Δ Confluence) at the timepoint of measurement compared to the start of the treatment was quantified. (A-E) Confluence progression of MCF-7 cells in (A) DMSO 0.1 %, (B) 10 nM E2 as positive control for estrogen-mediated increase in cell proliferation, as well as (C) 1 μ M, (D) 100 nM and (E) 10 nM BPTMC. (F-H) Confluence progression of MCF-7 cells with 100 nM of the selective ER α antagonist fulvestrant to mitigate estrogenic effects: (F) DMSO 0.1 %, (G) 10 nM E2, and (H) 1 μ M BPTMC. (I) Δ Confluence at the end of the 72 h treatment period. Values represent mean \pm SD of three independent experiments, each performed in six technical replicates. Statistical significance was determined using one-way ANOVA, followed by Tukey's post-test, with * p < 0.05.

estrogen-induced proliferative response. Similarly, treatment with BPTMC at 100 nM and 1 μ M decreased the doubling time, whereas 10 nM BPTMC had no noticeable effect (Suppl. Figure 5). Importantly, co-treatment with fulvestrant reversed the doubling time of both E2- and BPTMC-treated cells to that of vehicle treated MCF-7 cells, confirming that the proliferative effects are mediated through ER α (Suppl. Figure 5). Taken together, these findings demonstrate that BPTMC promotes the proliferation of ER α -positive breast cancer cells at nanomolar concentrations, highlighting its potential estrogenic and pro-proliferative properties in hormone-sensitive tissues.

4. Discussion

The transactivation assays performed in HEK-293 cells expressing human ER α and a responsive luciferase reporter construct revealed that BPTMC is a significantly more potent ER α agonist than BPA. In contrast, BPSIP tended to activate ER α at 10 μ M, whereas cyclo-di-BADGE and TBBPS were inactive. None of the four compounds tested exhibited antagonistic activity towards ER α . Molecular modeling further showed

that BPTMC possessed higher lipophilicity, lower solubility, and a smaller dipole moment than BPA, indicating greater overall hydrophobicity. Consequently, binding of BPTMC to the ER α ligand-binding site released more $\Delta G_{\text{desolvation}}$, thereby strengthening ligand-receptor interactions. This enhanced interaction may be a possible explanation for the observed increase in ER α activation potency compared to BPA. Interestingly, BPTMC, but not BPA, was found to be a moderate inhibitor of the E2-inactivating 17 β -HSD2, which could lead to elevated local E2 concentrations, thus enhancing local ER α activation. Importantly, BPTMC strongly upregulated the expression of the ER α target genes *GREB1*, *TFF1*, and *PGR*, starting at nanomolar concentrations in the MCF-7 breast cancer cell line, which endogenously expresses ER α . The magnitude of gene induction was comparable to that elicited by 10 nM E2 and was completely abrogated by the ER α antagonist fulvestrant, confirming receptor-specific activation. Consistent with its transcriptional activity, BPTMC also promoted proliferation of MCF-7 cells at nanomolar concentrations, an effect similarly reversed by fulvestrant. Collectively, these findings identify BPTMC as potent ER α agonist, with additional inhibitory activity against the E2-inactivating enzyme 17 β -

HSD2. However, additional effects on other nuclear receptors and steroid metabolizing enzymes remain to be clarified.

The European Chemicals Agency (ECHA) has classified BPA as a “substance of very high concern” due to its endocrine-disrupting properties and reproductive toxicity (den Braver-Sewradj et al. 2020). In line with European Union regulatory frameworks, BPA and several of its derivatives have demonstrated endocrine-disrupting effects, which are of particular concern for sensitive populations such as pregnant and breastfeeding women, given the importance of estrogens in the development of reproductive and mammary gland systems (Beausoleil et al. 2018). A body of evidence has linked BPA exposure to a broad range of adverse health outcomes, including developmental, metabolic, reproductive, immunological, and oncogenic effects (Rochester, 2013). As a result, BPA has been subjected to increasingly stringent regulatory restrictions and partial bans. Nonetheless, the rapid introduction of functional BPA substitutes, often without robust toxicological evaluation, raises concerns of “regrettable substitutions”, wherein replacement compounds may pose equal or even greater health hazards. Although efforts to assess BPA alternatives have intensified and regulatory agencies, including the European Commission, have emphasized the need for comprehensive and transparent safety assessments (European Commission 2024), substantial knowledge gaps remain. The ongoing phase-out of BPA and related compounds, driven by new restrictions within the European Union, has accelerated the demand for functionally equivalent replacements (European Commission 2024). Among the emerging alternatives, BPTMC has already been identified as a candidate BPA substitute and prioritized for toxicological evaluation (van den Brand et al., 2024). Like BPA, BPTMC has been detected in both environmental and human biomonitoring samples (Franko et al. 2024; Lucarini et al. 2020; Sauer et al. 2021). Its detection in 55 % of baby bottles sold in Canada as of 2021 underscores its growing commercial relevance and potential for widespread exposure (Siddique et al. 2021). Despite its increasing use, BPTMC remains poorly studied in terms of endocrine activity and systemic toxicity. Other BPA replacements have raised similar concerns. BPSIP, introduced as a dye-developing agent in thermal paper, has limited biomonitoring data available. However, it was detected in thermal paper receipts in Switzerland at concentrations ranging from 3.4 to 13.2 mg/g (Goldinger et al. 2015). Notably, BPSIP was more frequently detected in blood samples than BPA and bisphenol S (BPS), suggesting greater persistence and potential for systemic exposure (Thayer et al. 2016). Cyclo-di-BADGE, a degradation product or polymerization byproduct that is not covalently bound within polymer networks, is prone to migration into food. In this context, cyclo-di-BADGE and other BADGE derivatives have been identified in canned food samples (Bustos et al. 2023). TBBPS is a flame retardant and has been found globally in various environmental media. While it is used as an alternative to tetrabromobisphenol A (TBBPA), scientific opinion suggests that TBBPS and its derivatives may not be entirely environment friendly. Studies indicate that TBBPS can cause various toxic effects, including neurotoxicity, hepatotoxicity, and cytotoxicity (Xu et al. 2024). Despite the increasing prevalence of these BPA analogues in consumer products and the environment, toxicological data remain limited (Pelch et al. 2019).

Given the structural similarities among bisphenol derivatives (Fig. 1) and their demonstrated bioactivities, only a limited number of studies have assessed their estrogenic activity. In this study, we confirmed the ER α agonistic effect of BPA. This aligns with previous findings demonstrating that BPA can compete with the natural ligand E2 for binding to the ligand-binding pocket (LBP) of ER α (Delfosse et al. 2012; Paris et al. 2002). Although the estrogenic potency of BPA is significantly lower than that of E2, it has been shown to elicit estrogenic responses *in vivo* through direct interaction with ER α (Matthews et al. 2001). While extensive data exist for BPA and its most common replacements (e.g., BPS, bisphenol F (BPF), bisphenol AF (BPAF)), information on many other bisphenol analogues and substitution candidates remains sparse, inconsistent, or entirely lacking (Pelch et al. 2019). Earlier studies have

assessed estrogenic effects of some BPA replacement compounds using MCF-7 breast cancer cells. For most of the studied compounds, ER α -mediated estrogenic effects were observed either alone or in mixtures. Notably, BPAF, bisphenol B (BPB), BPF, BPS, bisphenol Z (BPZ), tetrachlorobisphenol A (TCBPA), and tetramethylbisphenol A (TMBPA) exhibited estrogenic activity at the mid- and high nanomolar range, comparable to or exceeding that of BPA (Bockers et al. 2020; Kim et al. 2017; Kitamura et al. 2005; Matteo et al. 2025; Zhao et al. 2019). In contrast, bisphenol P (BPP), diphenylmethane (DPM), diphenylpropane (DPP), tetramethylbisphenol F (TMBPF), Pergafast 201, and 2,4'-BPS showed no detectable activation of ER α -mediated signaling (Kitamura et al. 2005; Matteo et al. 2025). Studies of bisphenolic mixtures suggested that their estrogenic effects are largely additive (Matteo et al. 2025).

In this context, our study provides novel evidence that BPTMC exhibits potent estrogenic activity in human cell-based systems, with robust ER α activation observed at mid-nanomolar concentrations. To the best of our knowledge, this is the first study demonstrating such pronounced estrogenic effects of BPTMC in physiologically relevant models. This is especially relevant considering that the less potent BPA has been restricted for its endocrine-disrupting properties. This finding contrasts earlier reports that either showed minimal ER α activation by BPTMC or identified only weak estrogenic effects in the micromolar range (Dvorakova et al. 2018; Elke Eilebrecht et al., 2019; Keminer et al. 2020). These discrepancies may reflect differences in assay sensitivity, cell model relevance, or the use of non-physiological exposure levels. Our results underscore the importance of employing human-derived cell systems and realistic concentration ranges when evaluating endocrine activity. Our *in silico* analysis of BPA and BPTMC binding shows that hydrophobicity and pose rigidity affected binding affinity to ER α . This suggests that broader virtual screening of bisphenol properties might be beneficial to identify BPA derivatives with estrogenic activity. ER α is widely distributed throughout the body, in a variety of tissues. ER α is expressed primarily in the uterus, mammary glands, liver, kidneys, skeletal muscles, and heart (Couse and Korach, 1999). It is also expressed in the epididymis, thyroid, bones, adrenals (Brandenberger et al. 1997; Pau et al. 1998), and certain regions of the brain (Couse and Korach, 1999), as well as granulosa cells (dependent on follicle stage), ovarian surface epithelium, and theca cells (Saunders et al. 2000). E2 plays a critical role in many physiological processes in both females and males. These include normal growth, development, and cell-type-specific gene regulation in tissues of the reproductive tract, central nervous system, and skeleton (Chen et al. 2022). Since BPTMC is a potent ER α agonist, it may interfere with estrogen-regulated physiological processes, potentially leading to adverse effects in both central and peripheral systems.

In addition to its estrogenic properties, BPTMC has been reported to exert strong anti-progestagenic and weak anti-androgenic effects in transgenic human U2-OS osteosarcoma cells (Sauer et al. 2021). Moreover, comparative cytotoxicity assessments have revealed that BPTMC induces marked cell death in various reproductive cell models, including C18-4 spermatogonial cells (7.9 μ M), steroidogenic MA-10 (Leydig) cells (4.5 μ M), and KGN (granulosa) cells (14.7 μ M), often with greater potency than BPA or other bisphenol analogues (Rajkumar et al. 2021). In contrast, we did not observe any cytotoxicity in MCF-7 cells (15,000 cells/well) exposed for 24 h to BPTMC at concentrations up to 3 μ M. The apparent discrepancy (particularly for Leydig cells) may be explained by experimental differences, including cell type, lower cell density (2,000–6,000 cells/well), serum composition, and longer exposure duration (48 h), likely accounting for the cytotoxic responses observed by Rajkumar et al. (2021). Notably, yeast-based reporter assays have detected ER α agonism by BPTMC only at relatively high concentrations (e.g., EC₁₀ = 4 μ M and EC₅₀ = 3 μ M) (Dvorakova et al. 2018; van Leeuwen et al. 2019), suggesting lower sensitivity of these systems and further underscoring the need of mammalian cell-based models to evaluate endocrine activity. Assuming a scenario where BPTMC

eventually reaches tissue concentrations comparable to pre-regulation BPA levels, relevant testing concentrations would likely fall in the low nanomolar range. For example, BPA levels in breast adipose tissue of breast cancer patients have been measured at 4.2 ng/g, translating to \sim 17 nM when adjusted for tissue density and molecular weight (Baldwin et al. 2010; Keshavarz-Maleki et al. 2021). Tissue concentrations of BPA are generally higher in lipid-rich compartments and can reach up to 50-fold higher in fetal tissues, reflecting its lipophilic nature and potential for bioaccumulation (Geens et al. 2012; Nahar et al. 2015). Although human exposure data for BPTMC remain very limited, its higher lipophilicity compared with BPA suggests a potentially greater tissue accumulation (Zhao et al. 2023). The only reported plasma concentration of BPTMC is 0.014 μ g/L (\sim 45 pM) with reported values ranging from non-detectable levels up to 0.207 μ g/L (\sim 0.7 nM) (Hu et al. 2025), while urinary levels were below detection (Lucarini et al. 2020). Based on BPA data, where adipose tissue concentrations are approximately 4–5 times higher than plasma, and considering that BPA can reach up to 50-fold higher levels in certain fetal tissues, a similar concentration might be expected for BPTMC. Applying the 4–5x tissue-to-plasma ratio to the highest reported human plasma level (0.7 nM) yields estimated BPTMC concentrations of \sim 3–4 nM in lipid-rich tissues, and potentially up to \sim 30–35 nM in fetal compartments. Given that the *in vitro* EC₅₀ for BPTMC is 87 nM, these estimated tissue levels fall within an order of magnitude of the biologically active concentrations determined in our MCF-7 assays, where ER α -mediated gene activation and proliferation were observed. Although industrial production of BPTMC seems to be increasing considerably, no data are currently available on annual production of BPTMC or its concentrations in animals although it has been detected in environmental samples (Sauer et al. 2021). This highlights the urgent need for systematic biomonitoring in both human and environmental matrices for BPTMC.

For other BPA derivatives, the available data remain incomplete or inconclusive. For example, BPSIP has shown no detectable ER α activity in human assays (Keminer et al. 2020), while TBBPS, despite computational predictions of ER α binding, exhibited such high cytotoxicity that functional evaluation was not feasible (Cao et al. 2017). Cyclo-di-BADGE, to our knowledge, has not yet been tested for ER α modulation and did not show robust agonistic or antagonistic activity against ER α in our study.

The estrogenic effect of BPA appears to be receptor-mediated, as it did not inhibit 17 β -HSD2 that inactivates E2 and its inhibitory effects on 17 β -HSD1 (IC₅₀ $>$ 100 μ M, (Chen et al. 2023)) requires concentrations far exceeding those found in biomonitoring studies, suggesting limited toxicological relevance. However, structurally related analogues may exhibit much higher potencies for enzyme inhibition. For instance, TBBPA inhibits 17 β -HSD1 at nanomolar concentrations, representing a 100-fold higher potency than BPA (Chen et al. 2023). We corroborated these findings experimentally, observing nanomolar IC₅₀ values of TBBPA against 17 β -HSD1 (data not shown). While several BPA derivatives have been shown to inhibit steroidogenic enzymes such as 3 β -HSD2 or 17 β -HSD1 (Chen et al. 2023; Chen et al. 2024), such inhibitory effects could not be observed for cyclo-di-BADGE, TBBPS, BPSIP, or BPTMC at concentrations up to 10 μ M in the present study. Nevertheless, this emphasizes the importance to include steroid-metabolizing enzymes in the analysis of endocrine-disrupting substances.

Recent evidence from a study using isolated granulosa cells suggested that BPTMC may influence steroid hormone production by altering gene expression (mRNA levels) of *HSD3B2* and modulating the secretion of progesterone and E2 (Iskandarani et al., 2025). Animal studies face limitations in translating findings to humans due to inter-species differences in metabolic pathways, as exemplified by species-specific variations in 17 β -HSD1 inhibition by BPA (Ye et al. 2011). *In vitro* systems are useful alternatives but face their own challenges, such as the lack of metabolic capacity. Future improvements may include co-culture models incorporating metabolizing cells or the use of

hepatic S9 fractions to simulate biotransformation.

17 β -HSD2 catalyzes the oxidation of active 17 β -hydroxy sex steroids into their less active or inactive 17-keto forms, including the conversion of E2 to E1, and testosterone to androstanedione, thereby exerting a protective function in peripheral tissues by preventing excessive local activity of these potent hormones (Moghrabi et al. 1997; Tsachaki and Odermatt, 2019). Data on the modulation of 17 β -HSD2 by BPA and its substitutes remain scarce. Our results indicate that BPA does not inhibit 17 β -HSD2 and has a very low affinity to interact with the catalytic site. The two BPA analogs TBBPS and BPSIP partially inhibited 17 β -HSD2 activity at 10 μ M, suggesting a moderate interaction with the enzyme. BPTMC inhibited 17 β -HSD2 activity with an IC₅₀ of approximately 5 μ M, exceeding the concentrations found in environmental and human samples and raising questions regarding the toxicological relevance of this observation. By inhibiting 17 β -HSD2, BPTMC could promote elevated local concentrations of E2 in peripheral tissues, which could enhance estrogen-dependent effects or pathologies, particularly in hormone-sensitive tissues. However, because ER α agonistic activity of BPTMC occurs at nanomolar concentrations, while 17 β -HSD2 inhibition requires micromolar levels, significant enzyme inhibition is unlikely at concentrations relevant for ER α activation. Thus, the primary mechanism underlying the endocrine-disrupting potential of BPTMC includes most likely direct activation of ER α . Nonetheless, at higher exposure levels, 17 β -HSD2 inhibition may still contribute to the overall endocrine-disrupting profile of BPTMC, potentially in an additive or synergistic manner, especially under complex exposure scenarios involving mixtures of BPA analogues.

In conclusion, our findings establish BPTMC as a potent ER α agonist, capable of inducing both transcriptional activity and mitogenic responses at low, nanomolar concentrations. This potency raises significant concerns regarding its potential as an EDC, particularly given its increasing use as a BPA substitute in consumer products. The observed biological activity of BPTMC at relevant concentrations suggests potential interference with hormone-regulated processes, including those involved in reproductive and developmental health, as well as hormone-responsive breast cancer, as indicated by its effects in MCF-7 cells, or endometriosis and endometrial cancer (not addressed in this study). Given these findings, BPTMC warrants further attention in toxicological evaluations and regulatory frameworks. Its emerging presence in products such as baby bottles, combined with the lack of comprehensive exposure data, highlights the urgency of further research into its pharmacokinetics, metabolism, and long-term health effects. Future studies on BPTMC should prioritize comprehensive toxicokinetic investigations, including the application of physiologically-based toxicokinetic (PBTK) models to better predict internal exposures across different populations and life stages. In parallel, research should also examine tissue-specific biological responses, potential cumulative or synergistic effects with other EDCs, and identify particularly vulnerable windows of susceptibility, such as prenatal and early postnatal development, when exposure may have long-lasting or irreversible consequences. In the broader context of chemical substitution policy, BPTMC exemplifies the risks associated with insufficiently tested BPA alternatives. These results underscore the critical need for a rigorous, mechanism-based screening strategy including comprehensive *in silico* characterization to evaluate replacement compounds before widespread adoption. Identifying structurally diverse, well-characterized, and biologically inert alternatives should be a top priority, not only to protect public health but also to ensure that regulatory efforts do not inadvertently perpetuate or exacerbate the very risks they aim to mitigate.

CRediT authorship contribution statement

Manuel Kley: Writing – review & editing, Visualization. **Jamal Bouitbir:** Writing – review & editing, Writing – original draft, Supervision, Resources, Project administration, Funding acquisition, Conceptualization. **Alex Odermatt:** Writing – review & editing, Writing

– original draft, Supervision, Resources, Project administration, Funding acquisition, Conceptualization. **Marie-Christin Jäger:** Writing – review & editing, Visualization. **Martin Smiesko:** Writing – original draft, Supervision, Software. **Rianne E. van Diest:** Writing – original draft, Investigation, Formal analysis, Data curation. **Friedrich L. Joos:** Writing – review & editing, Writing – original draft, Visualization, Validation, Investigation, Formal analysis, Data curation.

Declaration of Competing Interest

The authors declare the following financial interests/personal relationships that may be considered potential competing interests: Alex Odermatt and Jamal Bouitbir report financial support from the Swiss Centre for Applied Human Toxicology. The other authors declare that they have no competing financial interests or personal relationships that could have influenced the work reported in this paper.

Acknowledgement

This work was supported by the Swiss Centre for Applied Human Toxicology (SCAHT) to AO (No. SCAHT-GL 21–06) and JB (No. RV-AV 23–01). We thank Anne-Laure Demierre, Sabine Frey, and Orlando Mani from the Swiss Federal Office of Public Health (FOPH) for their valuable contributions to the discussion.

Appendix A. Supporting information

Supplementary data associated with this article can be found in the online version at [doi:10.1016/j.tox.2025.154393](https://doi.org/10.1016/j.tox.2025.154393).

Data availability

Data will be made available on request.

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